

SIMULTANEOUS REMOVAL OF DISSOLVED ORGANIC MATTER, Microcystis aeruginosa, AND MICROCYSTIN-LR BY PRE-OXIDATION AND COAGULATION-FLOCCULATION PROCESSES

REMOCIÓN SIMULTÁNEA DE MATERIA ORGÁNICA DISUELTA, Microcystis aeruginosa Y MICROCISTINA-LR POR LOS PROCESOS DE PREOXIDACIÓN Y COAGULACIÓN-FLOCULACIÓN

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Abstract

The aim of this work is to assess the effect of the presence of *Microcystis aeruginosa* (MA) and microcystin-LR (MC-LR) on the removal of dissolved organic matter (DOM) by coagulation-flocculation operated at optimal conditions. Also, the effect of pre-oxidation with NaOCl on the simultaneous removal of MA, MC-LR, and DOM was studied, using the same processes and conditions. The water used in the experimental tests was prepared with distilled water, strains of MA, microcystin-LR standards, and model compounds of DOM. All trials were carried out in a jar test equipment applying two designs: a fractional factorial design to identify those variables with significant influence, and a response surface to determine the optimal operational conditions for each of the evaluated processes. Results show that the simultaneous presence of MA and MC-LR reduce until 32.25% the DOM removal, because they were competitors for the coagulant. The main effect was associated to MA due to its physicochemical characteristics and its major concentration. The pre-oxidant dose did not have a significant influence on the removal of MA, MC-LR and DOM since the contact time used in the tests was very short.

Keywords: cyanobacteria, cyanotoxins, clarification, sodium hypochlorite, dissolved organic carbon, drinking water .

Resumen

En este estudio se evaluó el impacto de la presencia de *Microcystis aeruginosa* (MA) y de la microcistina-LR (MC-LR) sobre la remoción de la materia orgánica disuelta (MOD) por coagulación-floculación en condiciones óptimas. Adicionalmente, se determinó el efecto de la preoxidación con NaOCl en la remoción simultánea de MA, MC-LR y MOD, empleando esos mismos procesos y condiciones. En las pruebas se utilizó agua destilada dopada con una cepa de MA, con estándares de MC-LR y con compuestos modelo de la MOD. Los experimentos se realizaron en un equipo de prueba de jarras, utilizando un diseño factorial fraccionado para determinar los factores de influencia significativa, y uno de superficie de respuesta para estimar las condiciones óptimas de operación de los procesos evaluados. Los resultados mostraron que la presencia de MA y MC-LR disminuye 32.25% la remoción de la MOD por un efecto de competencia por el coagulante, asociado principalmente a la MA, ya que presentó una mayor concentración y características fisicoquímicas que la hacen más susceptible de removerse por coagulación. La dosis de preoxidante no tuvo influencia significativa en la remoción de los tres contaminantes, debido a que el tiempo de contacto utilizado en las pruebas fue bajo.

Palabras clave: cianobacterias, cianotoxinas, clarificación, hipoclorito de sodio, carbono orgánico disuelto, agua potable.

1 Introduction

Dissolved organic matter (DOM) includes all organic compounds with hydrophilic (HPI) and hydrophobic (HPO) characteristics that pass through a 0.45 μ m pore diameter membrane. These compounds

are generated through natural biological, chemical and photochemical reactions issued from the decomposition of plants and animals in aqueous medium, and they are present in all types of water bodies (Hendricks, 2006; MWH, 2012). Removal of DOM is important because it affects the water organoleptic quality (color, odor, and taste), is a precursor of disinfection by-products, and it is the

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most abundant fraction of natural organic matter in water (up to 90% measured as dissolved organic carbon (DOC)) (Hendricks, 2006; Matilainen *et al.*, 2010; MWH, 2012). However, this fraction is the most difficult to remove by conventional potabilization processes due to its physicochemical characteristics: highly hydrophilicity or solubility, with negative electrical charge, a molecular weight between 500 and 3000 Da, and a composition integrated by 45-60% carbon, 4-5% hydrogen, 35-40% oxygen, and 1% nitrogen (Hendricks, 2006; Sillampaa, 2015). Furthermore, the removal of the DOM can be reduced by seasonal growths or blooms of cyanobacteria and its metabolites (AWWA, 2010; Zamyadi *et al.*, 2012).

Cyanobacteria are prokaryotic, autotrophic, photosynthetic organisms which are classified as Gram-negative bacteria. A bloom or exponential growth of cyanobacteria occurs when the conditions in a water body are adequate for its development (temperature, nutrients, and pH). Nevertheless, cyanobacterial blooms have increased their frequency in number and intensity due to climate change and human activities (AWWA, 2010; May-Cua et al., 2019; Navarro-Peraza et al., 2017; Quiblier et al., 2013). The main problem of blooms is the excessive production of toxic metabolites, known as cyanotoxins, which are part of the DOM and represent a health risk, even in trace concentrations. Drinking water treatment plants that use coagulationflocculation do not have the capacity to efficiently remove these contaminants during blooms. As a result, the effluent does not comply the required organoleptic and toxicological water quality, the filtration runs of sand filters decrease, the demand of coagulants, flocculants, adsorbents and oxidants increases, as well as the operation and maintenance costs of water treatment plants (AWWA, 2010).

MC-LR is a cyanotoxin produced by most of the cyanobacterial species, which is very stable in water, one of the most toxic and it has been related with the highest number of intoxication reports in human beings (Stewart *et al.*, 2006; Westrick *et al.*, 2010; Pantelic *et al.*, 2013; Merel *et al.*, 2013). In Mexico, there are some reports of the presence of extracellular MC-LR (dissolved in water) with values above the limit (1 μ g·L⁻¹) recommended by the World Health Organization (WHO) (Alillo *et al.*, 2014; Vasconcelos *et al.*, 2010). As consequence, the MC-LR maximum concentration is the most regulated in international standards for drinking water. Mexico recently began to establish a standard (COFEMER, 2017). Due to its high abundance and physicochemical properties (hydrophobicity, negative charge, and solubility), MC-LR competes with the rest of the DOM compounds to be removed through conventional potabilization methods. For these reasons, MC-LR and cyanobacterium MA, one of the main producer of this cyanotoxin, were selected to be evaluated in this work (Šejnohová, 2008).

Coagulation-flocculation (CF) is one of the most commonly conventional process used in water treatment facilities due to its low cost and because it is easy to operate. Only in Mexico, 24% of the installed drinking water treatment plants utilize it, thus this is equivalent to 69% of the drinking water production (SEMARNAT, 2016). Coagulationflocculation has individual removal efficiencies of cyanobacteria ranging from 70 to 90%, cyanotoxins from 0 to 39%, and DOM (quantified as DOC) from 10 to 60% (Lambert et al., 1996; Hendricks, 2006; Merel et al., 2013b). However, the simultaneous removal of these three pollutants, and the effect of cyanobacteria and their cyanotoxins on the DOM removal by coagulation-flocculation have been scarcely studied. Thus, the main objective is to improve the CF efficiency to remove DOM once the three compounds are present. In particular it is priority to increase the removal of cyanobacteria because they cause operational problems to water treatment plants and they have a high sanitary importance (Sillampaa, 2015). For this purpose, one suggests the combination of CF with other processes such as pre-oxidation, which has been reported as a coagulant aid, easy to apply and has a low costs (Matilainen et al. 2010; Sillampaa, 2015).

In this work, NaOCl was applied as a coagulant aid in the simultaneous removal of these pollutants using the optimal operating conditions for a potabilization train, which is composed of the pre-oxidation-coagulation-flocculation-sedimentation (PCFS) processes.

2 Methodology

The methodology was divided into three main stages: the first one consists in the preparation of the water samples to be treated; the second one involves the elaboration of two experimental designs (fractional factorial to determine the factors of significant influence and response surface to obtain the optimal conditions for the maximum simultaneous removal of DOM, MA and MC-LR); and the last stage includes laboratory tests performed to determine the effect of the presence of both MA and MC-LR on the removal of DOM and the application of preoxidation with NaOCl to evaluate its impact on the simultaneous removal of the three pollutants under optimal operating conditions for the treatment train proposed in this study.

2.1 Materials and reagents

Polluted samples, prepared in a type II water (resistivity 15 M Ω .cm), were used in all experimental tests carried out in this work. The preparation of these samples consisted in adding different amounts of one MA strain, MC-LR standards, and seven representative model compounds of the different DOM fractions present in surface waters (Table 1). DOM was quantified as Dissolved Organic Carbon or DOC (mg/L) in a TOC-LCSH Shimadzu analyzer. The strain of Microcystis aeruginosa was obtained from the bacteriology laboratory of the FES-Iztacala, UNAM; its culture was carried out under dynamic conditions at a temperature of 28 ± 0.5 °C in a 14h photoperiod and it was quantified in a Neubauer chamber (cell mL^{-1}). MC-LR standard (Cayman Chemical) was used applied and its concentration was quantified by HPLC-PDA with a previous concentration from solid phase extraction using STRATA-X cartridges. To provide alkalinity and turbidity to the water samples, sodium bicarbonate (NaHCO₃) (J.T. Baker) and bentonite (SIGMA-ALDRICH) were used, respectively. Aluminum sulfate (Al₂ (SO₄)₃·18H₂O) (J.T. Baker) was used as coagulant and sodium hypochlorite (NaOCl) as a pre-oxidant. The pH was adjusted using sulfuric acid (H₂SO₄) and sodium hydroxide (NaOH), both of reactive grade.

2.2 Stage 1: Preparation of water samples for treatment

In each experimental laboratory test, 2 L of synthetic water sample were used; the samples were prepared in distilled water by adding: 1) 1.5 g of sodium bicarbonate to obtain an alkalinity of 750 mg L^{-1} as CaCO₃; 2) DOM compounds as presented in Table 1, in amounts according to the experimental design, in all cases the DOM initial concentration was 5 mg DOC L^{-1} ; 3) the MA strain; 4) the MC-LR standard; 5) bentonite to obtain an initial turbidity of 5 UTN, and 6) sulfuric acid or sodium hydroxide to adjust the initial pH. The DOC concentration used in the tests was the average value reported for most of the drinking water supply surface sources. The turbidity was the average value reported for most of water surface supplies during the dry season. The alkalinity value was selected to avoid pH variations due to the coagulation-flocculation process (Hendricks, 2006; MWH, 2012).

2.3 Stage 2: Experimental designs

After preparing the water samples, trials of preoxidation-coagulation-flocculation-sedimentation were performed in a jar test equipment (JARTESTER PB 700 PHIPPS and BIRD), based on two experimental designs, which were established using the STATGRAPHICS software and selecting as response variables the removal efficiency of DOM, MA, and MC-LR. The duration of each process was established based on the times applied in "Los Berros" (Ortíz, 2014) and "Madín" (Félix, 2010) water treatment plants, and also on the results reported by Cerón *et al.* (2016) (Table 2).

Fraction	% DOC	Model compound
Hydrophobic acid (HPOA)	47	Fulvic and humic acids
Hydrophobic basic (HPOB)	2	Aniline
Hydrophobic neutral (HPON)	3.7	Gallic acid
Hydrophilic acid (HPIA)	16.3	Citric acid
Hydrophilic basic (HPIB)	3	Glutathione
Hydrophilic neutral (HPIN)	28	Polysucrose

Table 1. Typical composition of DOM fractions as DOC and representative model compounds.

Source: Imai et al., (2001); Kennedy et al., (2005); Zularisam et al., (2006/2007)

	processe	es.
Parameter	Time	Gradient
Pre-oxidation	10 s	600 s ⁻¹ (273 rpm)
Coagulation	1 s	600 s ⁻¹ (273 rpm)
Flocculation	5 min	100 s ⁻¹ (83 rpm)
	5 min	42 s ⁻¹ (45 rpm)
	10 min	25 s ⁻¹ (30 rpm)
Sedimentation	30 min	-

Table 2. Experimental conditions of P-C-F-S

The first experimental design used in this work was a fractional factorial type to evaluate the effect of six factors and to determine those of significant influence on the three response variables (removal of DOC, MA, and MC-RL). The formula to calculate the number of experiments is $2^{k-p} + n_c$, where: 2 is a constant that indicates the levels of the factors (-1 and +1 in coded values), k is the number of factors, p indicates the experimental design fraction to be selected for performing the tests, and n_c is the number of replications of the central point experiment. A quarter of experiments of the original design was selected with three replications of the central point, thus a total of 19 experimental tests were performed with the preoxidant. With a quarter of this design a resolution IV is obtained, in other words the main effects of the factors can be estimated and, as a consequence, the determination of those parameters of significant influence can be made. In addition, the replicates in the central point allowed to evaluate the existence of curvature, which indicates the need of a higher order model (Gutiérrez and de la Vara, 2008). The factors evaluated in this experimental design were: initial pH, coagulant dose, pre-oxidant dose, initial concentration of MA, initial concentration of MC-LR, and the percentage ratio between hydrophobic and hydrophilic water characteristics (HPO:HPI).

Table 3 shows the codified and natural values of the evaluated factors, as well as the references and the criteria applied for their selection.

For *M. aeruginosa* (MA), the cell density reports for algae consortiums detected in the Valle de Bravo reservoir reached maximums of 1×10^5 cell mL⁻¹ on October 2000 and April 2001 and for later on June 2010, it was 14×10^4 cell mL⁻¹ (Gaytan *et al.*, 2011; Alillo *et al.*, 2014). The value selected for this study was of 1×10^5 cell mL⁻¹.

The second experimental design was based on the response surface, using the Box-Behnken type $(2^k + 2k + n_c)$, to evaluate the influence of four factors. For this design, three replications of the central point test were performed, thus 27 experimental tests were carried out. The experimental results were analyzed to obtain a statistical model that was used to calculate the optimal value of each factor to allow the maximum simultaneous removal of contaminants. This design was of second order, thus it is able to determine the optimum point within the limits of the experimental space established by high (1) and low (-1) levels of the factorial design (Gutiérrez and de la Vara, 2008).

Factors	Unity		Level		Criterion	Reference
	-	-1	1	0		
[MC - LR]o	μg/L	1.5	3.25	5	Maximum dissolved concentration reported in a body of water to supply the largest water treatment plant in Mexico	Alillo <i>et al.</i> (2014)
[Coagulant]o	mg Al/L	3.6	8.55	13.5	Typical aluminum dose for initial DOC concentrations until 10 mg/L	Cerón <i>et al</i> . (2016)
pH	_	5	6.5	8	Typical pH in surface water	Cerón et al. (2016)
[<i>MA</i>] <i>o</i>	cell/mL	20 000	60 000	100 000	Minimum and maximum concentrations found in a body of surface water	WHO (1999)
HPO:HPI	%:%	80:20	50:50	20:80	Typical variations of water composition	Cerón <i>et al</i> . (2016)
[NaOCl]o	mg/L	0.25	1.125	2	Typical doses of NaOCl that do not cause damage to the cell wall of cyanobacteria	Zamyadi (2012)

Table 3. Real and coded values of the factors evaluated by fractional factorial design.

2.4 Stage 3: Laboratory water treatment tests

Jar tests were performed in duplicate applying the calculated optimal conditions to determine the effect of pre-oxidation with NaOCl in the simultaneous removal of MA, MC-LR, and DOM and the influence of the presence of MA and MC-LR on the DOM removal.

3 Results and discussion

3.1 Factors of significant influence (factorial experimental design)

An analysis of variance (ANOVA) with a 95% confidence level of the results of the factorial design tests (a quarter of the experimental design) was performed. The number of data values of this design were sufficient to identify factors of significant influence on the response variables (removal of DOM, MA, and MC-LR) (Gutiérrez and de la Vara, 2008).

3.1.1 Factors of significant influence on the DOM removal

The Pareto chart of the DOM removal results (quantified as DOC) is presented in Figure 1, which shows that the factors of significant influence were: coagulant dose, HPO:HPI ratio, and pH. Due to the coagulant dose showed a positive effect on the process, thus a dose increase will produce a rise of the DOM removal efficiency. This factor was the one with the greatest significant influence in these tests, probably due to its capacity to destabilize the particles charge and ability to favor the formation of complexes with contaminants with their subsequent removal from water (Hendricks, 2006). The effect of the HPO:HPI ratio was negative, and thus the DOM removal efficiency will augment as the concentration of the hydrophobic organic matter increases; these results agree with data reported in literature (Zhang et al., 2012; Sillampaa, 2015), which are explained on terms of the HPO organic matter composition. This DOM fraction is mainly composed of humic substances with high molecular weight, which consists of several phenolic structures, conjugated double bonds, and carboxyl groups that are efficiently removed using aluminum salts. The pH showed a negative significant influence, and thus a lowering



Fig. 2. Factors of significant influence in the removal of MA by P-C-F-S.

of the pH value represents a greater DOM removal. This effect is associated to pH determines the aluminum species, the coagulation mechanism, the speciation of NaOCl, and the ionization of the different functional groups of MON, MA, and MC-LR (Rivasseau et al., 1998; Hendricks, 2006; Merel et al., 2010; Zhang et al., 2012); therefore, at acidic pH values, the removal can be favored due to the predominance of Al polymeric species, which have higher density of positive charge compared to the monomeric species formed at higher pH. Also, in acid conditions the hypochlorous acid (HOCl) concentration increases (which is the most reactive species of the NaOCl hydrolysis), the stability of MA decreases, and the dissociation of DOM and MC-LR compounds increments (Rivasseau et al., 1998; Hadjoudja et al., 2010; Meneses, 2011; Cerón et al., 2016).

3.1.2 Factors of significant influence on the MA removal

Figure 2 shows the Pareto chart for the MA removal, which illustrates that the MA initial concentration was the only factor of significant and positive influence; consequently, the MA removal depends on its own initial concentration, and thus the efficiency will increase with an augmentation of this factor. As shown in Figure 2, the magnitude of the influence of this factor was significantly higher than the one of others. This means that the natural value of the upper level of the factor has been significantly low to observe a significant effect of the others (Gutiérrez and de la Vara, 2008), and therefore it is recommended to increase the value of this factor. However, it is difficult to detect concentrations of MA higher than the cell density proposed, considering the reports of Alillo et al. (2014), Gaytan et al. (2011), and WHO (1999); thus, it was chosen that the value of this factor remained constant for the optimization tests,



Fig. 3. Factors of significant influence in the removal of MC-LR by P-C-F-S.

which was the maximum concentration used in the first experimental design.

3.1.3 Factors of significant influence on the MC-LR removal

The statistical analysis (Pareto chart) of the MC-LR removal data is presented in Figure 3, which shows that the only significant and positive influence factor was the initial concentration of MC-LR. As occurred for the MA, this indicates that the initial value of the upper level used for this factor in the experimental design was excessively low and thus did not allow observing an effect of the others. However, it was selected to keep constant the initial MC-LR concentration at the highest value used in this experimental design, because it is the maximum dissolved concentration detected in a surface water used as supply source in the largest Mexican water treatment plant (Alillo *et al.*, 2014).

All factors of significant influence (coagulant dose, pH, and the HPO:HPI ratio) on the process were evaluated in the optimization tests. Also, even if the pre-oxidant dose in the studied interval did not show a significant influence on the MC-LR removal, this factor was evaluated since there is controversy about its effect as a coagulant aid (Meneses, 2011; MWH, 2012).

3.2 Optimization of the factors values of significant influence for the maximum simultaneous removal of pollutants (Box-Behnken design)

The value of coagulant dose, the HPO:HPI ratio, pH, and the pre-oxidant dose was optimized in this experimental stage. The test results were used to perform an analysis of variance with a confidence level of 95%, using the STATGRAPHICS software. This analysis fulfilled all the assumptions (normality,

constant variance, and independence) to validate the results obtained for each response variable (DOM, MA, and MC-LR); in addition, none of the residuals exceeded ± 2 sigmas to be classified as an atypical data since a value outside this range indicates a significant deviation from normality (Gutiérrez and de la Vara, 2008). The lack of fit was not statistically significant, indicating that the second order model obtained in this work describes adequately the behavior of the three response variables (p-value = 0.2548 for DOM, p-value = 0.4858 for MA, and p-value = 0.2712 for MC-LR).

The value of the determination coefficient (R^2) was appropriate to be used as a goodness-of-fit test, because the number of factors evaluated was less than five (Gutiérrez and de la Vara, 2008). The models and R^2 values are shown in Table 4; in all cases, the values of the determination coefficient were greater than 70%, and therefore the models are suitable for prediction purposes (Gutiérrez and de la Vara, 2008).

The desirability function was used to find a simultaneous removal optimum of the three pollutants. This function is applied if there is more than one response variable and allows to determine the point where all variables show simultaneously a satisfactory value (Gutiérrez and de la Vara, 2008). Desirability can have values between 0 (unacceptable) and 1 (the maximum desirable value). Table 5 shows the optimal value estimated by the desirability function for each of the evaluated factors to obtain the maximum simultaneous removal of the three pollutants; at these conditions the model calculated a desirability value of 0.8.

The experimental results obtained using the optimal conditions are shown in Table 6. When the removal values calculated by the model and the experimental results were compared, DOC showed the smallest difference. The experimental removal efficiencies obtained for MA and MC-LR were significantly lower than the values calculated by the model; this may be related to the significant experimental error reported for the analytical techniques used for the quantification of these pollutants, which could reach up to 30% (Bastidas, 2013).

These results, normalized by milligram applied of aluminum (Table 7), show that the depletion efficiencies of DOM and MA are within the ranges reported in studies conducted to remove separately both contaminants (Merel *et al.*, 2013b; Hendricks, 2006).

Model	R ² (%)
$\mathbf{DOM} = 2.91 + 0.09 * A + 0.78 * B - 0.19 * C + 0.07 * D - 0.45 * A^2 - 0.05 * A * B + 0.03 * A * C - 0.24 * B + 0.03 * A * C + 0.03 * C + 0.03 * A * C + 0.03 * C + 0.$	70
$A * D - 0.60 * B^{2} + 0.58 * B * C + 0.46 * B * D - 0.02 * C^{2} + 0.41 * C * D - 0.47 * D^{2}$	
$\mathbf{MA} = 74444.3 - 15277.8 * A + 33888.9 * B - 5833.3 * C - 18333.3 * D - 2777.7 * A^2 + 833.5 * A * B + 12500.0 * B + 12500.$	84
$C = 4166.5 * A * D = 10694.3 * B^2 = 2500.0 * B * C + 31666.7 * B * D = 25277.7 * C^2 = 27500.0 * C * D + 21805.7 * D^2$	76
$\mathbf{MC-LR} = 1.54 + 0.41 * A + 0.30 * B - 0.80 * C - 1.51 * D + 0.83 * A2 + 0.79 * A * B + 0.54 * A * C - 0.48 * 0.54 * 0.54 * A * C - 0.48 * 0.54 * 0.5$	76
$A * D = 0.51 * B^2 = 0.99 * B * C + 0.34 * B * D + 1.02 * C^2 = 0.40 * C * D + 0.39 * D^2$	

Table 4. Model of each response variable and its determination coefficient.

A = pre-oxidant, B = coagulant, C = HPO:HPI relation and D = pH

Table 5. Optimal values estimated for each factor based on the function desirability (value = 0.8).

Factor	Estimated optimal value
Pre-oxidant dose (mg NaOCl/L)	0.44
Coagulant dose (mg Al/L)	12.23
HPO:HPI relation (%:%)	80:20
pH	7.35

Table 6. Simultaneous removal of pollutants by applying optimal operating conditions of the P-C-F-S processes.

Response variable	Model	Experimental
DOM (mg DOC/L)	2.56	2.59 ± 0.35
MA (cell/mL)	123397	90000 ± 5774
MC-LR (μ g/L)	4.23	3.68 ± 0.02

However, the elimination of MC-LR was 35% higher than the average obtained in other studies (Lambert *et al.*, 1996). These results showed that in this study these three contaminants (DOM, MC-LR, and MA) could be efficiently removed simultaneously.

3.3 Effect of MA and MC-LR on the DOM removal by pre-oxidation and coagulation-flocculation

Figure 4 shows the amounts of DOM removed per mg of aluminum for three scenarios: with only the DOM present in water, with DOM and MC-LR simultaneously present, and finally when the three pollutants were simultaneously present.

The DOM removal at optimum operating conditions, using pre-oxidation, was 0.31 ± 0.002 mg DOC mg Al⁻¹ in the absence of MA and MC-LR, and 0.21 ± 0.01 mg DOC mg Al⁻¹, in the presence of both contaminants, which represents an efficiency depletion of 32.25%. The presence of MA produced a diminution of 0.08 mg DOC mg Al⁻¹, which is a decrease of 25.8% of the efficiency, and MC-LR only diminished 0.02 mg DOC mg Al⁻¹ (6.45%). This efficiency decrease can be associated to the competition of both cyanobacteria and their cyanotoxins with DOM for the consumption of NaOCl (pre-oxidant or coagulation aid) and aluminum sulfate (coagulant).

This competition is related to the initial concentration of each pollutant as well as to its physicochemical characteristics (size, solubility, surface charge, chemical composition), which can make them more susceptible to an attack with the pre-oxidant or to be removed by the coagulant.

Table 7. Simultaneous normalized and percentage removal of pollutants.

Response variable	Removal		Individual removal efficiencies reported %
	Normalized	%	
DOM	7,359 ± 385 mg DOC/mg Al	89 ± 7	10-60
MA	0.21 ± 0.01 cell/mg Al	49.45 ± 4.9	70-90
MC-LR	$0.30 \pm 0.01 \mu\text{g/mg}$ Al	73.57 ± 0.42	0-39



Fig. 4. Effect of the MA and the MC-LR on the removal of DOM (mg DOC/mg Al) in the clarified, applying the optimal conditions of operation of the processes of P-C-F-S.

On the one hand, the pre-oxidant can interact with the conjugated aromatic compounds of DOM, with some groups of MC-LR (3-amino-9-methoxy-10phenyl-2,6,8-trimethyl-deca-4,6-group-dienoic acid), and can also affect the electrostatic stability and growth of MA (Ho et al., 2006; Henderson et al., 2008; Shen et al., 2011; Ma et al., 2012). On the other hand, coagulation is more effective to remove colloids with sizes ranging from 0.001 to 10 μ m due to their negative charge, high hydrophobicity and molecular weight, such as humic and fulvic acids, which are removed after destabilization or neutralization of their charge by the coagulant. In contrast, coagulation is less effective in removing particles with no charge, hydrophilic character and low molecular weight, such as alkyls and polysaccharides, which can only be removed after formation of complexes or by adsorption on flocs (Hendricks, 2006; Gernaout et al., 2010). Considering this information and that MA is a colloidal particle (4-6 μ m) with negative surface charge and present in higher concentration than the one of MC-LR, thus it can be more easily removed by coagulation. For this reason, cyanobacteria had a greater impact on the decrease of the DOM removal. On the contrary, due to the small molecular size (0.003 μ m), low surface charge and concentration of the MC-LR, it is more difficult to be removed by coagulation, and consequently this pollutant impacted in a lesser degree the DOM removal (Hendricks, 2006; Aiken, 2014; Sillampaa, 2015).

3.4 Effect of the pre-oxidation process on the removal of MA, MC-LR, and DOM by coagulation-flocculation

In this stage a treatment test was carried out using the optimal operating conditions of the coagulationflocculation process, without addition of the pre-



Fig. 5. Effect of the pre-oxidant on the removal of DOM (mg DOC/mg Al), MA ($\times 10^4$ cell/mg Al) and MC-LR (μ g/mg Al).

and MC-LR removal. It was observed that there was no significant difference with or without application of NaOCl on the simultaneous removal efficiency of the three pollutants (Figure 5).

This result can be due to the low contact time (10 s) used in the test, which did not allow the development of none of the pre-oxidation mechanisms reported in literature (Henderson *et al.*, 2008; Ma *et al.*, 2012; MWH, 2012), such as transformation or alteration of the compounds, destabilization and inactivation of MA, and production of cellular exo-polymers, that serve as natural coagulants (Figure 6).

It is convenient to remember that this contact time is used in two of the main drinking water treatment plants installed in Mexico, "Los Berros" and "Madín" (Félix, 2010). Nevertheless, several laboratory studies only observed beneficial impacts of pre-oxidation using larger contact times (minutes). Lapsongpon et al. (2017) applied 3 mg L^{-1} of calcium hypochlorite in the pre-oxidation process and determined that it took 30 minutes to reach the maximum removal of MA by coagulation-flocculation. These authors mention that the removal was favored by the release of cellular exo-polymers. However, by applying longer contact times, a high generation of these compounds decreases the removal and increases the coagulant demand. They also observed that MA was inactivated with 2 mg L^{-1} and a contact time of 10 minutes, and this facilitated its removal and prevented its growth in subsequent processes. Ma et al. (2012) reported that after 20 minutes of pre-oxidation with a dose lower than 1 mg L^{-1} of NaOCl, the natural organic matter increased its molecular mass from 2524 to 2589; this improved 15% the DOC removal, and also the inactivation of MA was achieved. Nevertheless, it was found that with higher pre-oxidant doses, the release of cellular exo-polymers increases, and as



Fig. 6. The possible mechanisms of pre-oxidants and coagulant on simultaneous removal of DOM, MA, and MC-LR.

a result, the coagulant demand increased to obtain the same removal efficiency of pollutants.

Conclusions

In this study, *Microcystis aeruginosa* (MA), microcystin-LR (MC-LR) and dissolved organic matter (DOM) were simultaneously removed efficiently by applying a water treatment train with coagulation-flocculation(CF) as main process under optimal operating conditions. The removal efficiencies of the DOM and MA were within the ranges reported in literature, whereas the removal of MC-LR was 35% higher than the highest values reported.

DOM removal using CF decreased 32.25% in the presence of MA and MC-LR; the former caused the greatest depletion (25.8%) due to its colloidal characteristics and its higher concentration in raw water. In contrast, the MC-LR only decreased 6.45% the DOM removal efficiency, due to its significantly lower concentration (5 μ g L⁻¹) compared with the DOM content (5 mg DOC L⁻¹). In addition, MC-LR could compete with humic substances ($\leq 0.001 \mu$ m) for the coagulant dose due to its high surface charge and its low molecular size (0.003 μ m). The pre-oxidation process with NaOCl did not show a significant influence on the simultaneous removal of the three contaminants mainly for the low contact time used in the experiments. This short time did not permit to develop any of the reported mechanisms as enhancers of the CF process. This result is relevant because the contact time used in this work is similar to the values applied in drinking water treatment plants installed in Mexico. Therefore, it is recommended to use larger contact times than those used in such facilities, in order to have a significant beneficial effect in the simultaneous removal of MA, MC-LR, and DOM.

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Nomenclature

DOC	dissolved organic carbon
DOM	dissolved organic matter
HPI	hydrophilic
HPO	hydrophobic
MA	Microcystis aeruginosa
MC-LR	microcistina-LR
MOD	materia orgánica disuelta
NaOCl	sodium hydroxide
WHO	World Health Organization

References

- Aiken, G. (2014). Dissolved organic matter in aquatic systems. *Comprehensive Water Quality and Purification 1*, 205-220.
- Alillo, J., Gaytán, M., Martínez, V., and Ramírez, P. (2014). Microcystin-LR equivalents and their correlation with Anabaena spp. in the main

reservoir of a hydraulic system of central Mexico. *Inland Waters 4*, 327-336.

- American Water Works Association (AWWA). (2010). Algae: Source to Treatment; Manual of Water Supply Practices -M57, first edition. AWWA manual; M57, December, CO. EUA.
- О. (2013). Cell counting Bastidas. with neubauer chamber, basic hemocytometer usage. Celeromics, Technical Note-Neubauer Chamber Cell Counting, 1-6, (http://www.celeromics.com/en/resources/docs/ Articles/Cell-counting-Neubauer-chamber.pdf)
- Cerón, O., Martín, A., Rigas, F., Solís, M., and Ramírez, R. (2016). Optimization of the coagulation-flocculation process for the removal of natural organic matter fractions present in drinking water sources. *Water Quality Research Journal 51*, 153-166.
- Comisión Federal de Mejora Regulatoria (COFEMER). (2017). Proyecto de norma oficial mexicana PROY-NOM-127-SSA1-2017, agua para uso y consumo humano. Límites permisibles de la calidad del agua. No. Expediente: 02/0021/280617, pp. 1-170. (http://www.cofemersimir.gob.mx/expedientes/ 20393)
- Félix, T. (2010). Remoción de contaminantes emergentes en los sistemas de potabilización de la Ciudad de México, Tesis de Maestría del Programa de Maestría y Doctorado en Ingeniería, UNAM.
- Gaytan, M., Martinez, V., Oliva, M., Duran A., and Ramirez P. (2011). Temporal variation of phytoplankton from the tropical reservoir Valle de Bravo, Mexico. *Journal of Environmental Biology 32*, 117-26.
- Ghernaout, B., Ghernaout, D. y Saiba A. (2010). Algae and cyanotoxins removal by coagulation/flocculation: A review. Desalination and Water Treatment 20, 133-143.
- Gutiérrez, H. and de la Vara, R. (2008). *Análisis* y *Diseño de Experimentos*. Mc Graw Hill. Segunda Edición. México ISBN-13: 978-970-10-6526-6.
- Hadjoudja, S., Deluchat, V., and Baudu, M. (2010). Cell surface characterisation of *Microcystis*

aeruginosa and Chlorella vulgaris. Journal of Colloid and Interface Science 342, 293-299.

- Henderson, R., Parsons, S., and Jefferson, B. (2008). The impact of algal properties and pre-oxidation on solid-liquid separation of algae. *Water Research* 42, 1827-1845.
- Hendricks, D. (2006). *Water Treatment Unit Processes: Physical and Chemical*. Taylor and Francis Group, ISBN: 0-8247-0695-1.
- Ho, L., Onstad, G., Gunten, U., Pfeiffe, S., Craig, K., and Newcombe, G. (2006). Differences in the chlorine reactivity of four microcystin. *Water Research 40*, 1200-1209.
- Imai, A., Fukishima, T., Matsushige, K., and Hwan Kim, Y. (2001). Fractionation and characterization of dissolved organic matter in a shallow eutrophic lake, its inflowing rivers, and other organic matter sources. *Water Research* 35, 4019-4028.
- Kennedy, M., Chun, H., Quintanilla, V., Heijman, B., and Schippers, J. (2005). Natural organic matter (NOM) fouling of ultrafiltration membranes: fractionation of NOM in surface water and characterisation by LC-OCD. *Desalination 178*, 73-83.
- Lamber, T., Holmes, C., and Hrudey, S. (1996). Adsorption of microcystin-LR by activated carbon and removal in full scale water treatment. *Water Research 30*, 1411-1422.
- Lapsongpon, T., Leungprasert, S., and Yoshimura, C. (2017). Pre-chlorination contact time and the removal and control of *Microcystis aeruginosa* in coagulation. *Earth and Environmental Science* 67, 012011.
- Ma, M., Liu, R., Liu, H., Qu, J., and Jefferson, W. (2012). Effects and mechanisms of prechlorination on *Microcystis aeruginosa* removal by alum coagulation: Significance of the released intracellular organic matter. *Separation* and Purification Technology 86, 19-25.
- Matilainen, A., Vepsäläinen, M., and Sillanpää, M. (2010). Natural organic matter removal by coagulation during drinking water treatment: A review. Advances in Colloid and Interface Science 159, 189-197.

- May-Cua, E., Toledano-Thompson, T., Alzate-Gaviria, L. and Barahona-Pérez L. (2019). A cylindrical-conical photobioreactor and a sludge drying bed as an efficient system for cultivation of the green microalgae *Coelastrum sp.* and dry biomass recovery. *Revista Mexicana de Ingeniería Química 18*, 1-11.
- Meneses, R. (2011). Evaluación técnicoeconómica de los procesos de pre-oxidacióncoagulación-floculación-sedimentación utilizando hipoclorito de sodio y coagulantes poliméricos para la remoción de algas en agua de presa. Tesis de Licenciatura. Fes-Zaragoza UNAM. México.
- Merel, S., Villarín, M., Chung, K., and Snyder S. (2013). Spatial and thematic distribution of research on cyanotoxins. *Toxicon* 76, 118-131.
- Merel, S., Walker, D., Chicana, R., Snyder, S., Baurès, E., and Thomas, O. (2013b). State of knowledge and concerns on cyanobacterial blooms and cyanotoxins. *Environment International 59*, 303-327.
- Merel, S., Clement, M., and Thomas, O. (2010). State of the art on cyanotoxins in water and their behavior towards chlorine. *Toxicon* 55, 677-691.
- MWH. (2012). *Water Treatment Principles and Design*. Third edition. John Wiley and Sons Inc. New Jersey. EUA.
- Navarro-Peraza, R., Soto-León, S., Contreras-Andrade, I., Piña-Valdez, P., Viveros-García, T., Cuevas-Rodríguez E. and Nieves-Soto M. (2017). Effects of temperature and nitrogen limitation on growth kinetics, proximate composition and fatty acid profile of *Nannochloropsis sp. Revista Mexicana de Ingeniería Química 16*, 359-369.
- Ortiz, A. (2011). Aplicación de la coagulaciónfloculación acoplada a flotación con aire disuelto para la remoción de algas en un agua de presas. Tesis de maestría. UNAM. México.
- Pantelic, D., Svircev, Z., Simeunovic, J., Vidovic, M., and Trajkovic, I. (2013). Cyanotoxins: Characteristics, production and degradation routes in drinking water treatment with reference to the situation in Serbia. *Chemosphere 91*, 421-441.

- Quiblier, C., Wodd, S., Echenique, I., Heath, M., Villenueve, A., and Humbert, J. (2013). A review of current knowledge on toxic benthic freshwater cyanobacteria e Ecology, toxin production and risk management. *Water Research* 47, 5464-5479.
- Rivasseau, C., Martins, S., and Hennion, M., (1998). Determination of some physicochemical parameters of microcystins (cyanobacterial toxins) and trace level analysis in environmental samples using liquid chromatography. *Journal of Chromatography A 799*, 155-169.
- Secretaría de Medio Ambiente y Recursos Naturales (SEMARNAT). (2016). Inventario Nacional de Plantas Municipales de Potabilización y de Tratamiento de Aguas Residuales en Operación. Diciembre 2016. México. (http://files.conagua.gob.mx/conagua/ publicaciones/Publicaciones/SGAPDS-6-17.pdf).
- Šejnohová, L. (2008). Microcystis: New findings in Peptide Production, Taxonomy and Autecology. Institute of Botany, Czech Academy of Sciences, ISBN: 978-80-86188-27-0. Pruhonice, Czech Republic.
- Shen, Q., Zhu, J., Cheng, L., Zhang, J., Zhang, Z., and Xu, X. (2011). Enhanced algae removal by drinking water treatment of chlorination coupled with coagulation. *Desalination* 271, 236-240.
- Sillanpaa, M. (2015). Natural Organic Matter Characterization and Treatment Methods. Elsevier Inc. IWA Publishing. Oxford, UK. ISBN: 978-0-12-801503-2.
- Stewart, I., Webb, P., Schluter, P., and Shaw, G. (2006). Recreational and occupational field exposure to freshwater cyanobacteria - a review of anecdotal and case reports, epidemiological studies and the challenges for epidemiologic assessment. *Environmental Health. A Global* Access Science Source 2006, 6.
- Vasconcelos, V., Martins, A., Vale, M., Antunes, A., Azevedo, J., Welker, M., Lopez, O. y Montejano, G. (2010). First report on the occurrence of microcystins in planktonic cyanobacteria from Central Mexico. *Toxicon* 56, 425-431.

- Westrick, J., Szlag, D., Southwell, B., and Sinclair, J. (2010). A review of cyanobacteria and cyanotoxins removal/inactivation in drinking water treatment. *Anals of Bioanalytical Chemistry 397*, 1705-1714.
- World Health Oganization (WHO). (1999). Toxic Cyanobacteria in Water: A guide to Their Public Health Consequences, Monitoring and Management. Edited by Ingrid Chorus and Jamie Bartram. E & FN Spon. ISBN 0-419-23930-8. London.
- Zamyadi, A., Ho, L., Newcombe, G., Bustamante, H., and Prévost, M. (2012). Fate of toxic cyanobacterial cells and disinfection byproducts formation after chlorination. *Water Research* 46, 1524-1535.

- Zhang, X., Yang, Z., Wang, Y., Gao, B., and Yue, Q. (2012). The removal efficiency and reaction mechanism of aluminum coagulant on organic functional groups-carboxyl and hydroxyl. *Chemical Engineering Journal 211-212*, 186-194.
- Zularisam, A., Ismail, A., and Salim, R. (2006). Behaviours of natural organic matter in membrane filtration for surface water treatmenta review. *Desalination 194*, 211-231.
- Zularisam, A., Ismail, A., Salim, M., Sakinah, M., and Ozaki, H. (2007). The effects of natural organic matter (NOM) fractions on fouling characteristics and flux recovery of ultrafiltration membranes. *Desalination 212*, 191-208.