



## Biochar-packed biofilter for the treatment of gases produced by accumulated *Sargassum* waste

### Biofiltro relleno con biocarbón para el tratamiento de gases producidos por residuos acumulados de sargazo

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#### Abstract

A biofilter was built from the production of biochar obtained from *Sargassum* spp. (SKPH) as an environmentally friendly option for its application in problems related to the generation of gases such as H<sub>2</sub>S, CO<sub>2</sub>, and CH<sub>4</sub> that emanate during the decomposition process of accumulated *Sargassum* waste. The *Sargassum* was treated with KOH to obtain an activated biochar, estimating a surface area of 1319 m<sup>2</sup> g<sup>-1</sup> with an average pore size of 1.447 nm. This surface area value allows its use as a biofilter to control H<sub>2</sub>S, CO<sub>2</sub>, and CH<sub>4</sub> gases. The analysis of textural properties and structure of biochar were studied by XRD, FTIR, BET, and CHN-S techniques. The characterization results by X-ray diffraction and Raman spectroscopy showed a biochar with amorphous characteristics. In this study, the constructed biofilter was filled with SKPH biochar; this system was interconnected to the reactor containing *Sargassum* spp. (freshly collected from the beach) and to the portable gas meter. The results showed that sulfur, hydrogen, and nitrogen were significantly higher in the SKPH-LB sample. In general, biochar made it possible to demonstrate its efficient use to control H<sub>2</sub>S and CO<sub>2</sub> produced by the disposal of *Sargassum* waste, becoming a novel strategy to improve the elimination of toxic gases and reduce greenhouse gas emissions.

**Keywords:** biochar; biofilter; H<sub>2</sub>S; *Sargassum* spp., waste.

#### Resumen

Se construyó un biofiltro a partir de la producción de biocarbón obtenido del *Sargassum* spp. (SKPH) como una opción amigable con el medio ambiente para su aplicación en problemas relacionados con la generación de gases como el H<sub>2</sub>S, CO<sub>2</sub> y CH<sub>4</sub> que son emanados durante el proceso de descomposición de los residuos acumulados de sargazo. El sargazo fue tratado con KOH para la obtención de un biocarbón activado, estimándose un área superficial de 1319 m<sup>2</sup> g<sup>-1</sup>. Este valor de área superficial permite su uso como biofiltro para el control de los gases de H<sub>2</sub>S, CO<sub>2</sub> y CH<sub>4</sub>. Las técnicas XRD, FTIR, BET y CHN-S estudiaron las propiedades texturales y la estructura del biocarbón. Los resultados de la caracterización por XRD y espectroscopia Raman mostraron un biocarbón con características amorfas. En este estudio el biofiltro construido fue relleno con el biocarbón SKPH, este sistema se interconectó al reactor que contenía *Sargassum* spp. y al medidor de gas portátil. Los resultados mostraron que el azufre, el hidrógeno y el nitrógeno fueron significativamente mayores en la muestra SKPH-LB. En general, el uso del biocarbón demostró su uso eficiente para controlar el H<sub>2</sub>S y el CO<sub>2</sub> producidos por la descomposición de los desechos de sargazo, convirtiéndose en una novedosa estrategia para mejorar la eliminación de gases tóxicos y reducir las emisiones de gases invernadero.

**Palabras clave:** biocarbón, biofiltro, H<sub>2</sub>S, *Sargassum* spp, residuo.

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## 1 Introduction

In recent years, various materials such as iron-oxide-based materials (Feng *et al.*, 2020), zeolites (Rahmani *et al.*, 2023), polymeric membranes (Casadei *et al.*, 2020), and activated carbon (de Oliveira *et al.*, 2020) have been used to remove sulfurous compounds through adsorption. Hydrogen sulfide ( $H_2S$ ) is a colorless, extremely dangerous, corrosive, flammable gas heavier than air. It has a somewhat sweet and rotten egg-like odor and can be poisonous in high concentrations. At as low as 0.008 ppm, it causes an odor nuisance (Cox *et al.*, 2002).  $H_2S$  emissions are commonly found in volcanic eruptions, sulfur springs, fumes from underwater fissures, swamps, bodies of standing water, crude oil, and as a result of the organic decomposition of biomass (Cao *et al.*, 2023). Industries such as seafood processing, farming, leather markets, compost plants, and wastewater treatment (Olortiga-Asencios *et al.*, 2022) plants also produce  $H_2S$  (Letelier-Gordo *et al.*, 2020). A new factor contributing to the increase in  $H_2S$  production is the environmental problem caused by the decomposition of macroalgae in the Mexican Caribbean. *Sargassum* spp. has been arriving regularly since 2014, transforming the white sand and turquoise sea landscape into a picture of yellow and brown colors that is quite daunting for water and beach activities. This contingency has been recurrent, and between July and August 2015, an average of 9726 m<sup>3</sup> of seaweed accumulated per month per km of coastline was observed (van Tussenbroek *et al.*, 2017). This situation has led to various problems for the Quintana Roo coast in tourism, social, and environmental aspects. In the latter, the decrease in oxygen content, reduction in light, water quality degradation, beach erosion, etc. have caused mortality of near-shore seagrasses and fauna. *Sargassum* also affects daily life as rotten *Sargassum* produces  $H_2S$  when breathed in, and depending on the exposure time and concentration, it can cause serious health problems (Yaw Atiglo *et al.*, 2024).

High concentrations of  $H_2S$  in sediment have been linked to mortality events in seagrass (Koch *et al.*, 2022).  $H_2S$  is also responsible for serious environmental problems, such as acid precipitation and global warming. Other toxic compounds, such as carbon dioxide ( $CO_2$ ) and methane ( $CH_4$ ), which are part of greenhouse effect gases (GHG), can be stored in activated biochars that work as adsorbent materials (Gutiérrez-Bonilla *et al.*, 2022). Currently, biofilters have become a popular substitute for traditional air treatment procedures, including odor control and the removal of volatile organic compounds (VOC) (Permana *et al.*, 2022), as well as  $H_2S$  emissions. Biofilters are devices that transport a moist

contaminated fluid through a permeable filled area (Danila *et al.*, 2022). The effectiveness of biochar in a biofilter depends on numerous factors, including suitable packing materials, large surface area, specific surface area, pore volume, pore size distribution, good ion exchange capacity, high gas retention capacity, and improved selectivity (Das *et al.*, 2019). Recent research has shown that biochar can be beneficial when included as an adsorbent in biofilters (Yang *et al.*, 2023). Furthermore, biochar can be used alone or mixed with other substances as a filter medium in bioretention systems, water treatment, decontamination of urban stormwater, adsorption, etc. In research adsorption of  $CH_4$ ,  $H_2S$ , and  $CO_2$ , it was determined that the physicochemical properties of biochars were related to the adsorption performance (Sethupathi *et al.*, 2017). Adsorption has also been widely used for removing  $H_2S$ , using traditionally mesoporous materials with a larger surface area, which might enhance the adsorption potential (Izhar *et al.*, 2022).

Biochar, a porous carbon-rich material produced from biomass under oxygen-limited conditions, is now being developed as an efficient adsorbent that has been widely used for vapor analysis, including the adsorption of  $H_2S$  (Han *et al.*, 2020),  $CO_2$ , and  $CH_4$ . Moreover, biochar has a competitive advantage due to its low cost and can be obtained from any source of biomass. In recent years, various studies have been carried out that relate the type of biomass, the pyrolysis temperature, and the activation conditions using different atmospheres, and how these parameters affect the performance of biochar as  $CO_2$  and  $H_2S$  adsorbent materials, finding important differences in its behavior with each of these gases, related to both the porous structure and the surface chemistry. Several authors consider that H and S are stored in the pores of the biochar without being chemically linked to the biocarbon. That is, it can be assumed that  $H_2S$  adsorbs on active sites of the biochar, that is to say, the surface chemistry of biochar could be altered to anchorage specific functional groups to selectively  $H_2S$  adsorption (Elsayed *et al.*, 2009). A possible reaction mechanism during the  $H_2S$  adsorption onto biochar can be very complex with a diversity of sulfur products. Commonly, the reaction mechanics includes two steps: (1) physical adsorption in the presence of water,  $H_2S$  is adsorbed on the carbon surface, dissolved in water, and dissociated in an adsorbed state; (2) oxidation step. The adsorbed  $H_2S$  reacts with oxygen ( $O_2$ ) to form ending products of elemental sulfur (S0), sulfur dioxide ( $SO_2$ ), and sulfate ( $SO_4^{2-}$ ) (Huang *et al.*, 2022).

This study aims to present the findings of an experiment conducted in a laboratory using a biofilter that contained biochar SKPH. The biochar SKPH was derived from the remnants of *Sargassum* spp. and was

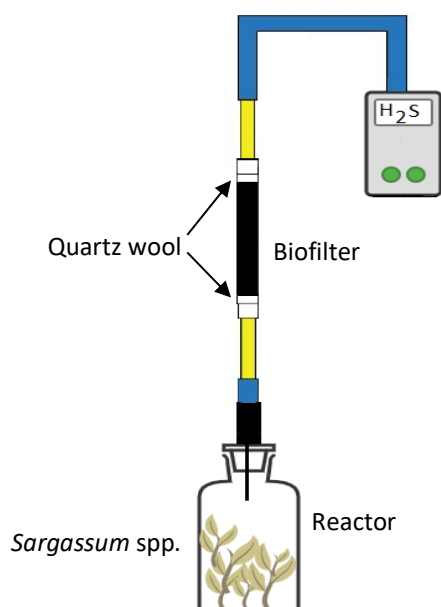


Figure 1 Diagram of the connection between the reactor and the biofilter.

used to eliminate greenhouse gases such as H<sub>2</sub>S, CO<sub>2</sub>, and CH<sub>4</sub>. We analyzed the physicochemical properties and adsorption capacities of the SKPH sample before and after it was connected to the reactor.

## 2 Materials and methods

### 2.1 Biofilter

In this study, biochar was produced from *Sargassum* spp. (sample SKPH) according to reported in previous work (Pérez-Salcedo *et al.*, 2019). The resulting biochar has an average pore size of approximately 1.4 nm and a measured surface area of around 1319 m<sup>2</sup> g<sup>-1</sup>. A biofilter was then constructed by filling a transparent glass tube (10 mm diameter, 125 mm height) with 400 mg of the SKPH biochar, and sealing both ends with 100 g of quartz wool. Silicone hose reducers (9 mm diameter, 15 mm height) were connected to both sealed ends, and a plastic tube (8 mm diameter, 60 mm long) was attached. This system was then connected to six reactors containing 10 g fresh weight (without any prior treatment), which were placed in 250 mL serological bottles that were sealed with a tert-butyl plug and secured with a metal ring. The bottles were then incubated at 35 °C for natural decomposition, with headspace gas samples taken individually every 48 hours for 16 days. CO<sub>2</sub> and H<sub>2</sub>S were measured using a Geotech 5000 gas analyzer portable equipment with three control units directly connected to the portable gas meter, and the other three reactors connected to the biofilter for 60

seconds. Figure 1 depicts the connection between the reactor, the sample of *Sargassum* spp, the biofilter, and the gas analyzer. SKPH-LB corresponds to the biochar sample that has been in contact with the gases generated by the decomposition of the biomass, that is to say, post-biofilter placement.

### 2.2 Characterization

Gas flow rates were measured and controlled using a Geotech Biogas 5000 gas analyzer in units of L min<sup>-1</sup>. The total pore volume, specific surface area, and pore size distributions were determined through N<sub>2</sub> physisorption tests using a Quantachrome Nova 2200e. The CHNS composition of biochars was examined using an automatic elemental analyzer, Thermoscientific Flash 2000. X-ray diffraction (XRD) patterns were obtained using Bruker D-2 Phaser with Cu-K $\alpha$  radiation from 10° to 80°. The optimal preparation condition was achieved in all the characterizations.

## 3 Results and discussion

### 3.1 Physical and chemical characteristics of biochar

#### 3.1.1 BET analysis

The biochar samples were degassed at 250 °C under vacuum for 5 hours to eliminate any impurities on their surface. The micropore BET assistant, which is included in the Novawin software, was employed to calculate the specific area ( $S_{BET}$ ) and the Brunauer-Emmett-Teller (BET) equation using relative pressures ranging from 0.05 to 0.3 P/P<sub>0</sub>. The pore volume and diameter of the biochar were determined using the Barrett-Joyner-Halenda (BJH) method. The estimation was based on the quantity of gas adsorbed at a relative pressure close to unity (0.98 P/P<sub>0</sub>). The pore size distribution graph was obtained through the novel Quenched Solid Density Functional Theory (QSDFT) method (Neimark *et al.*, 2009). Figure 2(a) displays the adsorption and desorption isotherms of the SKPH and SKPH-LB samples before and after being connected to the reactor containing *Sargassum* waste. The data revealed an II-type isotherm with an H<sub>4</sub> hysteresis loop for both SKPH and SKPH-LB, suggesting that these biochar samples consist of micro-mesoporous adsorbent materials (Bedin *et al.*, 2016). The pore structure parameters of biochar are a crucial factor that affects their adsorption properties. Figure 2(b) shows the pore size distribution of the biochars SKPH and SKPH-LB. Both samples show an undefined peak at around 1.4 nm which is attributed to the presence of micropores.

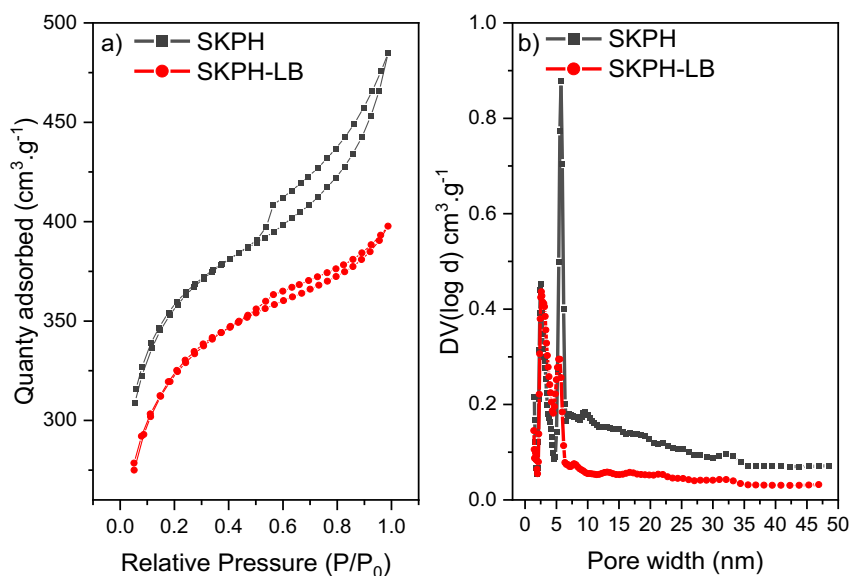


Figure 2 (a) Nitrogen adsorption-desorption isotherm of the samples SKPH and SKPH-LB, and (b) pore distribution.

Table 1 Pore structure parameters of the biochar samples.

Sample	BET surface area ( $\text{m}^2 \text{g}^{-1}$ )	Average pore size (nm)	Pore volume ( $\text{cm}^3 \text{g}^{-1}$ )
SKPH	1319	1.47	0.75
SKPH-LB	1191	1.41	0.61

Additionally, two increases in pore diameters can be seen at approximately 2.4 nm and 5.7 nm. These are associated with the formation of mesopores due to the activation process with KOH.

Table 1 provides a comparison of the average pore size, pore volume, and BET surface area for different biochar samples. The data in the table shows that the SKPH sample had a surface area of  $1319 \text{ m}^2 \text{g}^{-1}$ , while the SKPH-LB sample had a reduced surface area of  $1191 \text{ m}^2 \text{g}^{-1}$ . The reduction in surface area in the SKPH-LB sample could be due to the adsorption of nitrogen and sulfur trapped within the pores during the  $\text{H}_2\text{S}$  gas adsorption process, as revealed in the CHNS elemental analysis. This indicates that the micropores played a crucial role in the adsorption of  $\text{H}_2\text{S}$ . Furthermore, the pore diameter decreased slightly in the range of 5-5.6 nm, and according to the IUPAC classification, the biochar can be categorized as microporous. The adsorbed volume also decreased slightly in the 5-5.6 nm range, and to a lesser extent in the 1.40 to 1.50 nm range, indicating that both micropores and mesopores played a vital role in the  $\text{H}_2\text{S}$ ,  $\text{CO}_2$ , and  $\text{CH}_4$  adsorption process.

### 3.1.2 Elemental analysis CHN-S

The results of the elemental analysis of biochars are presented in Figure 3. The data shows that both samples, SKPH and SKPH-LB, have similar carbon contents, which are 74.42% and 76.37%, respectively. However, the sample SKPH-LB exhibits significantly

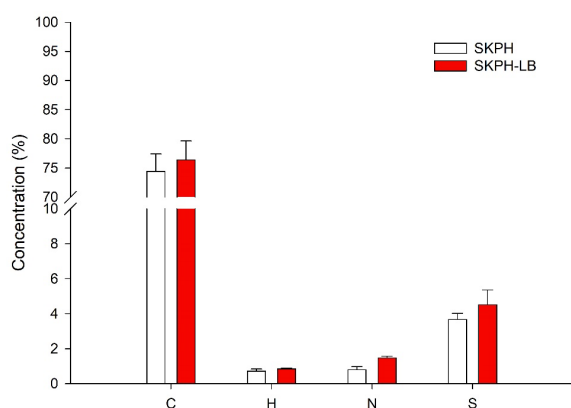


Figure 3 CHN-S analysis of the biochar SKPH and SKPH-LB.

higher concentrations of sulfur, hydrogen, and nitrogen. The sulfur concentration in SKPH-LB is 1.2 times higher than SKPH, while the hydrogen and nitrogen concentrations are 2.0 and 4.5 times higher, respectively. These increments represent increases of 107%, 355%, and 23%, respectively, compared to SKPH. Moreover, the increase in the amount of H, N, and S, in SKPH-LB suggests the retention of gases of  $\text{H}_2\text{S}$  and  $\text{CH}_4$  produced by the reactor containing *Sargassum* waste, thereby decreasing the accessible surface area ( $1191 \text{ m}^2 \text{g}^{-1}$ ). One of the outstanding attributes of biochar that may contribute to the biofilter performance is its ability to adsorb gases and this characteristic is related to the high surface area (Thompson *et al.*, 2020). Furthermore, the results indicate that carbon is the primary component of biochar. The C, H, N, and S composition analysis of biochar is summarized in Table 2.

Table 2 CHNS elemental analyses of biochar in weight percentage.

Biochar	C	H	N	S
SKPH	74.42	0.71	0.80	3.67
SKPH-LB	76.37	0.84	1.47	4.52

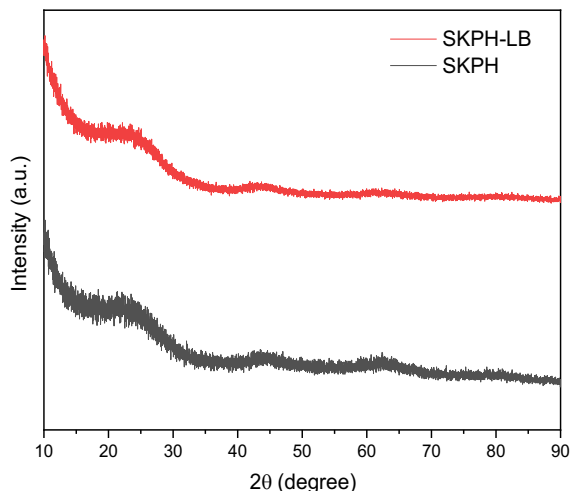


Figure 4 X-ray diffraction (XRD) patterns of SKPH and SKPH-LB.

### 3.1.3 X-ray powder diffraction (XRD)

In Figure 4, we can observe the amorphous carbon structures C (002) and C (100) for SKPH and SKPH-LB samples. The broad reflection between 20° and 25° indicates poor crystallinity without long-term crystalline order, in both samples, corresponding to the (002) plane of an ordered hexagonal graphitic carbon lattice. The peak located at 44.3° is associated with the plane (100), suggesting a more graphitic behavior (Wang *et al.*, 2020). For both samples, the diffractograms of the biochars are very similar and these confirm that it contained mainly the amorphous compounds due to the decomposition of biomass indicating an amorphous carbon structure, presence of C, and graphite with randomly oriented aromatic carbon sheets. It is worth mentioning that the crystalline nature of biochar is the key to its stability, and this property largely depends on the biomass source and carbon content. The presence of amorphous carbon is associated with random mixtures of thermally variable molecules, aromatic polycondensates, poorly structured graphene stacks anchored in the amorphous phase, and turbostratic materials composed of chaotic graphite crystallites (Tomczyk *et al.*, 2020).

### 3.1.4 Raman spectroscopy

Raman spectroscopy has been widely used to examine the carbon structure of both samples, pre (SKPH) and post-biofilter placement (SKPH-LB). In Figure 5, two peaks were observed at approximately 1322

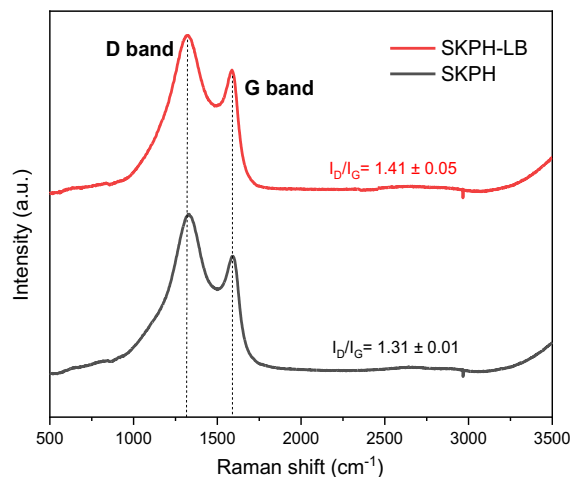


Figure 5. Raman spectra of the biochar SKPH and SKPH-LB.

and 1590  $\text{cm}^{-1}$ , corresponding to the D-band and G-band. The D-peak was attributed to  $\text{sp}^3$ -hybridized carbon defects, while the G-peak represented the vibrations of  $\text{sp}^2$ -hybridized graphitic carbon in a two-dimensional hexagonal lattice (Sun *et al.*, 2014). The results show a D and G peaks were very pronounced in both samples, and no significant differences were observed. It is possible to further analyze the diversity of defect degrees using the value of  $I_D/I_G$ . The ratio of the D peak and G peak's relative intensity ( $I_D/I_G$ ) indicated the structural defects and disorders in the biochars. The biochar SKPH-LB had a higher degree of disorderliness in the carbon structure, with  $I_D/I_G = 1.41 \pm 0.05$ , compared to sample SKPH, with  $I_D/I_G = 1.31 \pm 0.01$ . The increase in the defect degree of this biochar could be attributed to the introduction of N and S in the biochar, according to the results of the CHN-S elemental analysis.

### 3.2 Performance during the test

In Figure 6 (a) the line, indicated by black circles, shows the average values obtained from the direct  $\text{H}_2\text{S}$  emissions of the control group (SKPH). A concentration of 2,000 ppm was recorded during the first four days of sampling, and it reached the maximum value on day 10, with an average of 2,703 ppm. The lower line, indicated by empty circles, shows the average values obtained from the SKPH-LB sample passed through the biofilter. In SKPH-LB samples the maximum  $\text{H}_2\text{S}$  concentration was 120 ppm. This represents a greater than 95% decrease in the  $\text{H}_2\text{S}$  concentration concerning the control group emissions, which shows the efficiency of SKPH *Sargassum* biochar in removing this gas. In the literature there are other examples of the use of biochar, for example, biochar prepared from invasive algae *Sargassum* spp. was successfully employed as a sorbent for caffeine removal from water (Francoeur *et al.*, 2021).

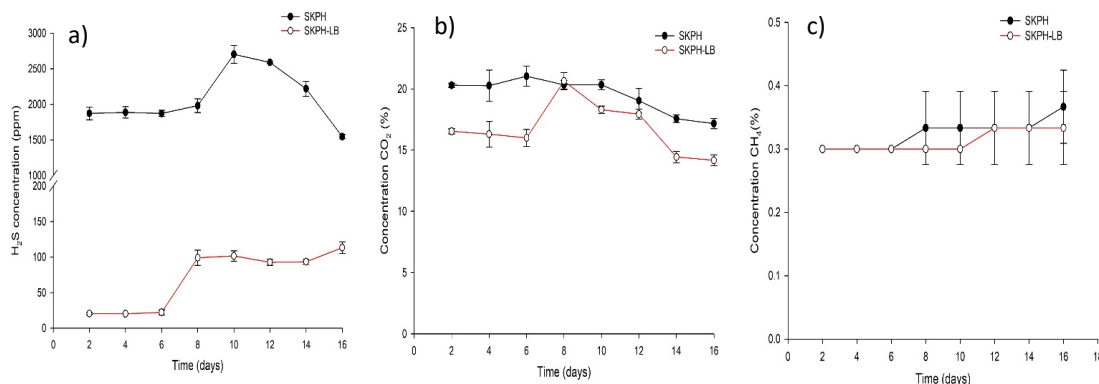


Figure 6 (a) Concentration of gases (a) H<sub>2</sub>S, (b) CO<sub>2</sub>, and (c) CH<sub>4</sub>.

In this same sense, Truong *et al.*, (2023) mentions the potential of biochar obtained from *Sargassum hemiphylum* to remove Cu(II) from aqueous solutions, leading to its high degree of applicability for water treatment. In addition, studies on *Sargassum* and *Enteromorpha* biochars indicate a high adsorption capacity which means good development potential in the field of H<sub>2</sub>S removal (Han *et al.*, 2020).

Figure 6 (b) shows the production of CO<sub>2</sub>, and as in the case of H<sub>2</sub>S, the values obtained from the direct measurement of the SKPH group are higher than those of the SKPH-LB sample. This suggests that biochar has the potential to absorb a significant portion of the CO<sub>2</sub> molecule (Zhang *et al.*, 2022). In Figure 6 (c), the detected methane concentrations were very low, with the highest concentration being only 0.36%. This low level of methane emission can be attributed to the fact that the *Sargassum* used in this experiment was recently collected and in this type of sample, mainly sulfate-reducing bacteria have been identified, which favors the production of H<sub>2</sub>S over methane (Hervé *et al.*, 2021). Furthermore, some authors have suggested that the high sulfur content in algae negatively affects methane production (Maneein *et al.*, 2021).

## Conclusions

An inexpensive and straightforward biofilter was tested for removing H<sub>2</sub>S, CO<sub>2</sub>, and CH<sub>4</sub> using accumulated *Sargassum* waste. The results showed that the biofilter made with biochar from *Sargassum* spp. can significantly absorb H<sub>2</sub>S and CO<sub>2</sub> gases. The nitrogen and sulfur content in the biochar increased after being exposed to reactor gases. To understand the physicochemical properties of the biochars, they were characterized by DRX, BET, and Raman analyses. The BET analysis revealed that the biochar's surface area decreased slightly after the H<sub>2</sub>S gas adsorption process, reaching a minimum value of 1208 m<sup>2</sup> g<sup>-1</sup> for sample SKPH-LB. This study opens an opportunity avenue for *Sargassum* waste, through the

use of biochar. Therefore, we consider biochar as an important candidate for further research in the emissions control of toxic gases.

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